

• 综 述 •

## 微生物除 Mn(II) 机制及影响因素研究进展

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**摘 要:** 锰是生物所必需的一种微量元素, 但工业技术发展以及矿产资源的开发导致大量的 Mn(II) 排放进入环境中对人体健康产生严重威胁。微生物修复技术可快速高效去除环境中的 Mn(II), 且无二次污染, 成为近年研究的热点。本文综述了除 Mn(II) 微生物的种类与分布及其除 Mn(II) 的机制, 总结了影响微生物除 Mn(II) 的因素, 并展望了除锰微生物的前景, 以期除锰微生物在锰污染水体中的高效应用提供理论参考。

**关键词:** 除锰微生物; 锰氧化酶; 影响因素; 吸附; 氧化

## The mechanism of microbial removal of Mn(II) and its influencing factors: a review

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**Abstract:** Manganese is an element essential for living organisms. Development of industrial technologies and exploitation of mineral resources have led to the release of large amount of Mn(II) into the environment, posing a serious threat to human health. Bioremediation can remove the Mn(II) from the environment rapidly and effectively without generating secondary pollution, thus received increasing attention. This review summarized the diversity and distribution of Mn(II) removal microorganisms and the associated mechanisms, followed by discussing the effect of environmental factors on microbial Mn(II) removal. Finally, the challenges and prospects for bioremediation of

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Mn(II) polluted wastewater were proposed.

**Keywords:** manganese removal microorganisms; manganese oxidase; influencing factors; adsorption; oxidation

锰元素是地壳中第二大过渡元素, 主要以 Mn(II)、Mn(III)、Mn(IV) 形式广泛存在自然界中<sup>[1-2]</sup>, 其中 Mn(II) 是植物体内超氧化物歧化酶的构成元素, 且参与了光合作用电子传递, 同时也是人体不可或缺的一种微量元素, 但已有研究发现摄入过量的 Mn(II) 会严重影响生物体发育, 甚至威胁人类生命<sup>[3-5]</sup>。随着工业技术的发展以及矿产资源的开发, 大量 Mn(II) 随着工矿企业污水排放和矿渣的随意堆放进入自然界中<sup>[6]</sup>, 对水体产生了诸多不利影响<sup>[7-8]</sup>。因此减少环境中 Mn(II) 的积累迫在眉睫。

传统的重金属修复方法如化学沉淀法、离子交换法、过滤法、膜分离法以及氧化还原法等具有成本高、处理效果不稳定、易引入 Cl<sub>2</sub>、KMnO<sub>4</sub>、O<sub>3</sub> 等新的化学物质造成二次污染的缺点<sup>[9-14]</sup>。生物修复技术具有成本低、操作简单、处理效果好且对环境无二次污染的优点, 不仅能快速有效地除去水体中的重金属且不会改变理化性质<sup>[15-18]</sup>。微生物可通过吸附或将 Mn(II) 氧化为不溶的锰氧化物沉淀进而降低环境中 Mn(II) 的毒性<sup>[19]</sup>, 生成的锰氧化物由于粒径小、比表面积大, 可吸附 Cd(II)、Pb(II)、Zn(II)、Cu(II) 等重金属甚至能降解一些如乙炔基雌二醇等环境有机污染物<sup>[20]</sup>, 实现对污染环境的二次修复。由于除锰微生物具有可高效去除 Mn(II), 成本低且无二次污染的优点, 成为环境 Mn(II) 污染的主要修复技术。不同种类微生物除 Mn(II) 过程存在差异, 且该过程会受到环境温度、pH 以及 Mn(II) 浓度的影响。

本文总结了除锰微生物的种类与分布, 论述

了微生物的除锰机制和主要影响因素, 并对微生物除锰技术的应用进行了展望, 以期除锰微生物的高效应用提供参考。

## 1 除锰微生物的种类及分布

除锰微生物 (主要为细菌和真菌) 在自然界中广泛存在, 大多数分布在含锰结核的土壤、沼泽、海洋和矿井等地<sup>[21]</sup>。表 1 列出了具代表性的除锰微生物。目前已分离到的除锰细菌种类较多, 包括假单胞菌 *Pseudomonas* spp.<sup>[22-23]</sup>、纤发菌 *Leptothrix* spp.<sup>[24]</sup>、节杆菌 *Arthrobacter* spp.、链霉菌 *Streptomyces* spp.、芽孢杆菌 *Bacillus* spp.<sup>[25-27]</sup>、紫外链霉菌 *Streptomyces violarius*、红球菌 *Rhodococcus* spp. 和分枝节杆菌 *Arthrobacter ramosus* spp. 等, 然而仅有少数假单胞菌、芽孢杆菌、土微菌和纤发菌是模式菌株<sup>[28-30]</sup>。目前已报道的除锰真菌较少, 主要有弯孢霉属 *Curvularia*、担子菌属 *Basidiomycetes*、链格孢属 *Alternaria*、轮生菌属 *Verticillium*、盾壳霉属 *Coniothyrium*、枝顶孢菌属 *Acremonium* 和青霉属 *Penicillium* 中的某些种<sup>[31-35]</sup>。虽然上述细菌和真菌均可除 Mn(II), 但由于它们体内锰氧化酶的种类不同, 因此在除锰方式上存在较大的差异。

## 2 微生物除锰机制

微生物可通过吸附和矿化作用达到除锰效果。生物吸附是微生物细胞壁组分和其分泌的胞外聚合物与 Mn(II) 结合形成络合物的过程; 生物矿化则是在微生物作用下使 Mn(II) 形成氧化物或氢氧化物等沉淀<sup>[53-55]</sup>。

表 1 代表性除锰微生物一览

Table 1 Representative manganese-removing microorganisms

Microorganisms	Sources	References
Bacteria		
<i>Pseudomonas putida</i> MnB1	Mn crust that accumulated in drinking water pipes in Trier, Germany	[22-23]
<i>Pseudomonas putida</i> GB-1	Green Bay nearly	[36-37]
<i>Pseudomonas aeruginosa</i> AT18	Petroleum-contaminated soil at the Petroleum Refinery "Hermanos Diaz" in Cuba	[38]
<i>Bacillus</i> sp. SG-1	The shallow marine sediment near San Diego, California	[39-40]
<i>Leptothrix discophora</i> SS-1	Swamp water	[24]
<i>Leptothrix discophora</i> SP-6	Surface film and flocculent material in the outflow reservoir of an artificial iron seep	[26]
<i>Lysinibacillus</i> sp. MK-1	A manganese mine in Hunan province, China, 27°42' north latitude, 111°58' east longitude	[40]
<i>Citrobacter freundii</i> FM-2	The biological activated carbon (BAC) filter column	[41]
<i>Arthrobacter</i> sp. HW-16	Manganese ore soil	[42]
<i>Streptomyces violarius</i> SBP1	Contaminated areas in Thailand	[43]
<i>Chryseobacterium</i> MSB-4	On the wall of the falling water aeration tank of the biological iron and manganese removal water plant	[44]
<i>Brevundimonas</i> MB-38	Water systems 1 in Virginia, USA	[45]
<i>Brevundimonas</i> AP5s2-K1a	The northwest shore of Ashumet Pond on Cape Cod, Massachusetts	[46]
<i>Stenotrophomonas</i> sp.	Acid mine drainage in Qibaoshan mine, Liuyang City, Hunan province, China.	[47]
<i>Flavobacterium</i> sp.	Acid mine drainage in Qibaoshan mine, Liuyang City, Hunan province, China.	[47]
<i>Brachybacterium</i> sp. Mn32	The Pacific seafloor near manganese cobalt nodules	[48]
<i>Duganella</i> AB14	Former Ronneburg uranium-mining district in Thuringia, Germany.	[31]
<i>Erythrobacter</i> sp.	A manganese mine in Hunan province, China, 27°42' north latitude, 111°58' east longitude.	[40]
<i>Thiobacillus thiophilus</i> sp.		
<i>Xanthomonas</i> sp.		
<i>Comamonas</i> sp.		
<i>Pedomicrobium</i> sp.	Drinking water distribution systems in south-east Queensland, Australia.	[24]
Fungi		
<i>Saccharomyces</i> sp.	The fermentor at a brewery located near Chennai, India.	[33,49]
<i>Acremonium strictum</i> DS1bioAY4a	Drainage system of acid mine in central Pennsylvania.	[50]
<i>Acremonium</i> sp. KR21-2	Manganese deposit surface in the Kikukawa river system (Shizuoka, Japan).	[51]
<i>Stagonospora</i> sp. SRC11sM3a	Drainage system of acid mine in central Pennsylvania.	[50]
<i>Phoma</i> sp. AP3s5J1a	A natural freshwater lake in the Ashumet Pond, MA.	[52]
<i>Pithomyces chartarum</i> DS1bioJ1b	Passive coal mine drainage treatment systems in Central Pennsylvania.	[52]
<i>Pyrenochaeta</i> sp. DS3sAY3a		
<i>Cephalosporium</i> sp.	In coarse-textured soil material underlying acid peat deposits in the humid regions of Newfoundland.	[31]

在生物矿化中,微生物锰氧化是生物除锰的主要方式。在生物锰氧化过程中微生物可通过直接或间接作用将 Mn(II) 氧化为 Mn(III) 或 Mn(IV) (图 1)。

## 2.1 生物吸附作用

生物吸附作用是微生物除锰的一种潜在机制,相比于细菌,真菌因具有大量细胞壁而拥有更强的吸附能力<sup>[56]</sup>。由于微生物细胞壁中存在胺、酰胺和羧酸等酸碱基团,当溶液 pH 升高时,这些官能团被脱质子化,导致细胞表面负电荷增加,从而促进对 Mn(II) 的吸收<sup>[32,38,57]</sup>。但也有研究发现 pH 只能影响微生物细胞壁对 Mn(II) 的吸附速率,并不能改变微生物对 Mn(II) 的吸附容量<sup>[38]</sup>。微生物吸附 Mn(II) 的化学官能团一般为 -OH、-NH、-C=O、-COOH 和 -C-O 等<sup>[32,47,58]</sup>。

## 2.2 微生物直接氧化 Mn(II)

微生物直接氧化 Mn(II) 的途径由其体内相关锰氧化酶介导。与锰氧化相关的酶有多种,以多铜氧化酶 (multicopper oxidases, MCO) 和锰过氧化物酶 (Mn peroxidase enzyme, MnP) 最为常见。

### 2.2.1 多铜氧化酶

多铜氧化酶是一类活性中心含 Cu(II) 的氧化酶,其以氧分子作为催化剂,电子通过酶内部多种类型的 Cu(II) 从底物转移到氧分子上,进而达到氧化有机化合物和各种金属的目的。随着

基因组学的发展,目前已有多种具有锰氧化活性的 MCOs 细菌被报道。进一步研究发现编码该酶的基因常以基因簇的形式存在,如短小芽孢杆菌 (*Bacillus pumilus*) WH4 的 *cotA* 操纵子<sup>[59]</sup>、*Bacillus* sp. PL-12、海洋芽孢杆菌 PL-12 的 *mnx* 操纵子<sup>[60]</sup>、恶臭假单胞菌 (*Pseudomonas putida*) MnB1 的 *ccm* 操纵子<sup>[22]</sup>、生盘纤发菌 (*Leptothrix discophora*) SS-1 的 *mofA*、*mofB* 和 *mofC* 操纵子<sup>[61]</sup>以及土微菌属 (*Pedomicrobium* sp.) ACM 3067 的 *moxA* 操纵子<sup>[62]</sup>等。

在 MCO 的锰氧化机制研究中, *mnx* 操纵子的调控研究最为详尽。该操纵子编码的 MCO 是由一个包含锰氧化活性中心的蛋白 MnxG 与两种辅助蛋白 MnE 和 MnF 形成的蛋白复合物。其调控 Mn(II) 的氧化步骤推测如下:首先 Mn(II) 在 Mnx 的作用下经水去质子化激活 Mn(II) 形成复合物 Mn(II)-OH, 然后通过桥连作用再次连接形成 Mn(II)(-OH)Mn(II), 此时被激活的 Mn(II) 失电子向 I 型铜转移<sup>[63]</sup>, 复合物中出现 Mn(III)<sup>[64]</sup>, 随后第二个 Mn(II) 结合并桥连第一个, 触发转移和 Mn(II) 移位, 两个周期缩合产生 Mn(III) ( $\mu$ -OH)<sub>2</sub>Mn(III), 最后通过歧化缩合反应生成 MnO<sub>2</sub>。目前认为该过程存在两种反应途径<sup>[63,65]</sup>, 分别是通过先歧化后缩合依次生成 Mn(OH)<sub>2</sub> 和 MnO<sub>2</sub> 或是先缩合后双歧化生成 MnO<sub>2</sub><sup>[66]</sup>(图 2)。

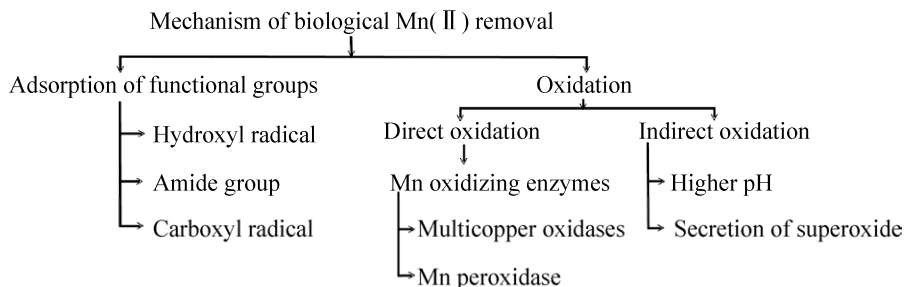


图 1 生物除锰机制

Figure 1 The mechanism of biological Mn(II) removal.

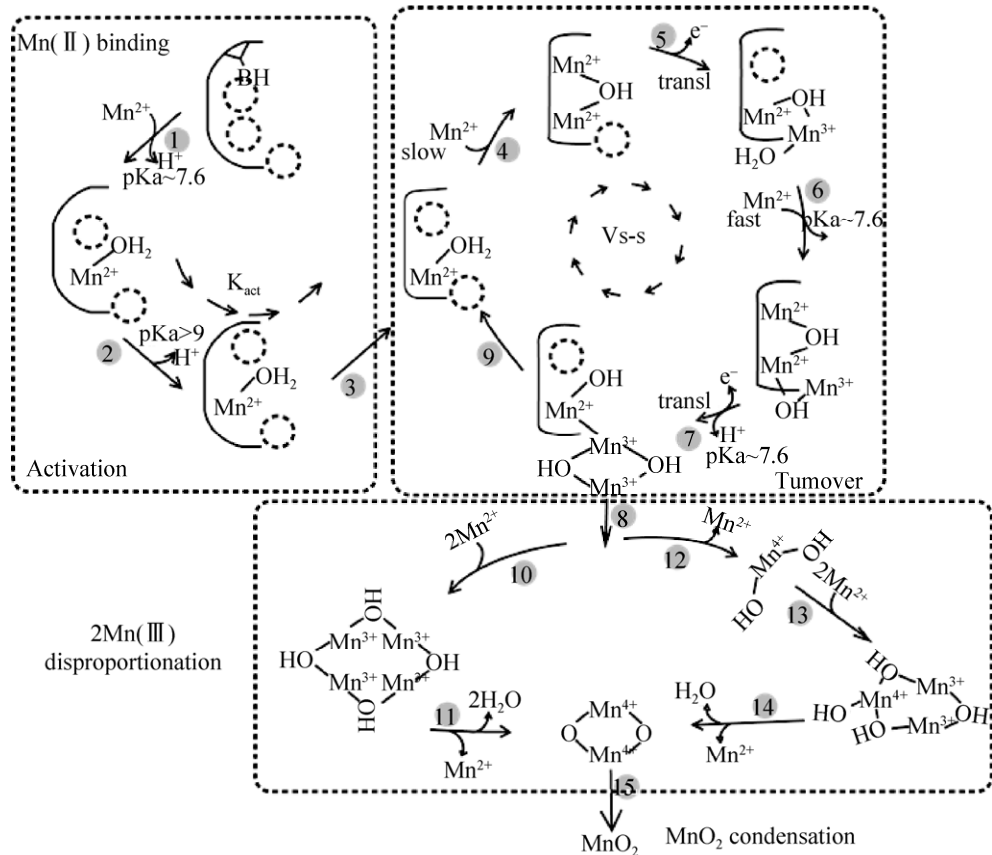


图2 MnX 催化氧化 Mn(II) 到 MnO<sub>2</sub> 的整体机制推断<sup>[63,66]</sup>

Figure 2 Proposed mechanism of MnX catalyzed oxidation of Mn(II) to MnO<sub>2</sub><sup>[63,66]</sup>.

### 2.2.2 锰过氧化物酶

锰过氧化物酶是一类含血红素的过氧化物酶,包括髓过氧化物酶(myeloperoxidases)、乳过氧化物酶(lactoperoxidases)和胞外锰过氧化物酶(Mn peroxidase enzyme, MnP)。其中髓过氧化物酶和乳过氧化物酶用过氧化氢作为氧化剂,目前仅在赤杆菌 *Erythrobacter* sp. SD21 和橙单胞菌 *Aurantimonas manganoxydans* SI85-9A1 中发现<sup>[67]</sup>。MnP 则主要存在于担子类真菌,如支顶胞属 *Acremonium* sp. KR21-2、禾顶囊壳小麦变种 *Gaeumannomyces graminis* var. *tritici* 和黄孢原毛平革菌 *Phanerochaete chrysosporium* DS2psM2a2 等菌株中<sup>[50-51,68]</sup>。

MnP 的催化循环过程与其他血红素过氧化

氢酶类似,但它更倾向于利用 Mn(II) 作为电子供体。在该过程中,微生物体内的 MnP 先被过氧化氢或有机物氧化后缺失两个电子形成 MnP 化合物 I,随后通过两个连续反应 MnP 恢复到最初形式,在该过程中形成 MnP 化合物 II。值得注意的是, Mn(II) 或其他电子供体存在下 MnP 化合物 I 均可转化为 MnP 化合物 II,但 MnP 化合物 II 只在 Mn(II) 存在时才能转化为初始形式的 MnP<sup>[69]</sup>。

### 2.3 微生物间接氧化 Mn(II)

虽然有些细菌如微藻菌 *Microcystis* sp. 和枯草芽孢杆菌 *Bacillus subtilis* 体内含有编码 MCO 或锰过氧化物酶的相关基因,但 Mn(II) 的氧化并未通过相关酶活介导,而是因培养液中 pH

和氧化还原电位 (Eh) 的改变实现<sup>[70-71]</sup>。此外, 有些细菌如寡养单胞菌 *Stenotrophomonas* sp.、球形芽孢杆菌 *Lysinibacillus* sp.、黏质沙雷氏杆菌 *Serratia marcescens* 和产酸克雷伯氏菌 *Klebsiella oxytoca* 基因组内未含有与 Mn(II) 氧化相关的基因, 但它们仍可通过提高培养液 pH 来促进 Mn(II) 的氧化<sup>[72]</sup>。这主要是由于它们在利用酵母粉和蛋白胨中的氨基酸进行生长时产生铵态氮  $\text{NH}_4^+$  积累在溶液中, 导致水体 pH 升高, 反应式如下:  $\text{NH}_3 + \text{H}_2\text{O} \rightleftharpoons \text{NH}_4^+ + \text{OH}^-$ 。

除提高溶液 pH 外, 有些微生物如玫瑰杆菌 (*Roseobacter* sp.) AzwK-3b 和子囊丝状真菌 (*Stilbella aciculosa*) 等在其细胞分化过程中还可通过胞外超氧化物对 Mn(II) 进行氧化, 氧化方程式为  $\text{Mn(II)} + \text{O}_2^- + 2\text{H}^+ \rightarrow \text{Mn(III)} + \text{H}_2\text{O}_2$ , 生成的过氧化氢可被微生物体内的过氧化氢酶分解产生氧气, 从而抑制其将生成的 Mn(III) 再还原为 Mn(II)<sup>[73-75]</sup>。

总之, 耐锰微生物除 Mn(II) 过程十分复杂, 有些微生物可同时通过吸附和氧化作用达到去除 Mn(II) 的效果。如 Tang 等<sup>[40]</sup>研究发现赖氨酸杆菌 *Lysinibacillus* sp. MK-1 对 Mn(II) 的总去除率为 94.67%, 其中生物氧化占 55.94%, 生物吸附占 36.23%; Zhang 等<sup>[19]</sup>通过研究也发现嗜水气单胞菌 (*Aeromonas hydrophila*) DS02 在前 24 h 内主要通过胞外吸附去除 Mn(II), 培养 144 h 后对初始浓度为 10 mmol/L 的 Mn(II) 的总去除率为 89.57%, 其中生物氧化占 49.55%。可见, 由于耐锰微生物种类复杂多样, 其体内催化的酶种类不尽相同, 因此对二价锰的去除机制也不同。

### 3 影响微生物除锰的因素

微生物在除锰过程中受到环境温度、初始 pH、Mn(II) 浓度等因素的影响。不同除锰微生物对外界条件的响应特征不一致, 目前倾向于筛

选适应性强且对锰去除率高的菌株。

#### 3.1 除锰微生物对温度的适应范围较广

除锰微生物对温度的敏感程度较低, 在 4–40 °C 之间可生长并具有去除 Mn(II) 的能力<sup>[10,76-77]</sup>。不同微生物种类对温度的耐受性不同, 导致其去除 Mn(II) 的最适温度也不一致。如 Zhao 等<sup>[77]</sup>发现短短芽孢杆菌 (*Brevibacillus brevis*) MO1 和类短短芽孢杆菌 (*Brevibacillus parabrevis*) MO2 在 37 °C 条件下对 50 mg/L Mn(II) 的去除效果最好, 在 30 d 内对 Mn(II) 的去除率分别为 82.6% 和 88.4%, 但当温度为 4 °C 时, 其对 Mn(II) 的去除能力均有所下降, 培养 30 d 后 Mn(II) 去除率分别为 50.7% 和 38.0%; 而赵炎报道的金黄杆菌 MSB-4 在 8–12 °C 条件下对 15 mg/L Mn(II) 的去除效果较好, 去除率高于 90%<sup>[44]</sup>。温度对除锰微生物的影响可能是通过影响微生物体内生长酶的活性来限制其生长进而影响其对锰的去除效果, 但也有研究发现在最适微生物生长的温度条件下菌株对锰的去除效果并未达到最大<sup>[44]</sup>, 说明温度对除锰微生物的影响是十分复杂的, 需进一步用酶学和分子生物学手段探讨温度对微生物除锰的影响。

#### 3.2 初始 pH 会影响除锰微生物吸附及氧化锰

提高溶液的初始 pH 可促进微生物细胞壁上吸附官能团 (胺、酰胺和羧酸等酸碱基团) 对锰离子的吸收以及微生物胞外聚合物对锰的氧化, 进而提高锰的去除效率<sup>[58,79]</sup>, 但当 pH 高出一定范围后, 溶液中过多的  $\text{OH}^-$  与锰离子结合形成氢氧化锰沉淀, 会阻碍微生物对锰离子的进一步吸收, 不同除锰微生物对 pH 的适应范围不一致。如节杆菌 (*Arthrobacter echigonensis*) MN1405 在初始 pH 为 6 时对锰离子的吸附量达到最大<sup>[80]</sup>, 而链要菌 (*Streptomyces spinoverrucosus*) NB-7 吸附锰离子的最佳条件是初始 pH 为 4<sup>[81]</sup>。此外,

初始 pH 还会影响微生物对锰的氧化,王文秀等<sup>[82]</sup>研究发现海洋锰氧化子囊真菌 (*Ascomycete*) PSA-107 h 在初始 pH 7 条件下对锰的氧化程度较高; Sasaki 等<sup>[83]</sup>在报道指出内楝内生真菌 (*Paraconiothyrium* sp.) 也存在类似现象。这可能是由于 pH 刺激微生物体内相关锰氧化酶的活性,进而对其氧化过程产生影响,但目前尚无相关研究。

### 3.3 环境中锰离子浓度会限制微生物除锰效率

不同微生物对环境中 Mn(II) 的耐受性不同,对 Mn(II) 的去除效果也不一致。*Aeromonas hydrophila* DS02 对初始浓度为 495 mg/L Mn(II) 的去除效果最好,去除率高达 89.57%,而当 Mn(II) 浓度升高到 990 mg/L 时,其对 Mn(II) 的去除率则降至 80% 以下<sup>[19]</sup>; *Lysinibacillus* sp. MK-1 对初始浓度为 55 mg/L Mn(II) 的去除率高达 98%,而当 Mn(II) 的初始浓度为 495 mg/L 时菌株 MK-1 对其几乎没有去除<sup>[84]</sup>。目前已筛选出的除锰微生物大部分可高效去除初始浓度约为 50 mg/L 的 Mn(II),去除率为 55%–100%<sup>[77,85]</sup>。由于近年来人类活动向自然界中排放的 Mn(II) 含量不断升高,导致水体中 Mn(II) 浓度增加,因此筛选可有效去除高浓度 Mn(II) 的微生物十分必要。

除上述 3 个因素外,其他因素如微生物的接种量<sup>[81]</sup>、作用时间<sup>[80]</sup>、环境中存在的其他离子(如铁离子、锌离子、铜离子、铵态氮等)<sup>[44]</sup>和有机物质(如抗生素)<sup>[22]</sup>也会影响除锰微生物对 Mn(II) 的去除效果。因此,在实际应用中,应依据现场情况合理选择菌源,并在可控范围内设置有效的工艺参数达到成本低、效果高的目的。

## 4 总结与展望

微生物可通过吸附或将 Mn(II) 氧化为不溶的锰氧化物沉淀从而降低环境中 Mn(II) 的

毒性<sup>[19]</sup>,微生物修复技术已成为锰污染治理中的有效手段。目前,可有效去除环境中 Mn(II) 的微生物被陆续报道,但已报道的微生物大多只可高效去除低浓度 (<50 mg/L) 的 Mn(II),且仅有少数菌株被用作模式菌株研究微生物除 Mn(II) 的机制。由于微生物种类不同,其对 Mn(II) 的去除机制也存在差异。本实验室前期从微生物菌剂中筛选获得 1 株可耐 5 000 mg/L Mn(II) 的细菌,分子生物学鉴定为肠杆菌属,而目前在肠杆菌属中尚未有耐锰微生物的报道,表明该菌是 1 株新型的耐锰细菌。进一步研究发现该菌在 2 d 内对 1 000 mg/L Mn(II) 的去除率高达 85.44%,且有 6.82% 的 Mn(II) 被氧化,说明该菌可通过氧化作用和其他作用高效去除 Mn(II),但具体机制尚未明确。此外,课题组前期研究发现金属离子会影响微生物对废水氮素的去除效果<sup>[86]</sup>,而筛选得到的耐锰菌株不仅可高效去除 Mn(II),对模拟废水中的铵态氮、硝态氮和亚硝态氮均有较好的去除效果。除 Mn(II) 外,氨态氮、硝态氮以及亚硝态氮也是废水中常见的污染物,大量的研究表明,微生物可通过自身代谢高效去除模拟废水中的氨态氮、硝态氮以及亚硝态氮<sup>[87-89]</sup>。少数研究发现实验室条件下,部分微生物可同步去除模拟废水中的 Mn(II) 和硝态氮或 Mn(II) 与氨态氮<sup>[78,90]</sup>。

在真实废水中, Mn(II)、氨态氮、硝态氮以及亚硝态氮通常同时存在,且微生物技术在实际应用过程中会受到众多因素(温度、pH、Mn(II) 浓度等)的影响。结合课题组前期研究成果以及目前锰污染水体处理面临的问题,未来需从以下几个方向展开研究:(1) 继续筛选可高效去除 Mn(II) 的微生物,丰富模式菌种类;(2) 通过电子顺磁共振(EPR)、对称性破损方法和密度泛函理论(BS-DFT) 结合分子生物学技术进一步深入探讨微生物除 Mn(II) 的具体机制;(3) 不断优化处理工艺技术,为除 Mn(II) 微生物营造

一个稳定的生长条件, 实现除 Mn(II) 微生物的高效应用; (4) 筛选可同步去除 Mn(II) 和不同形态氮素的微生物, 并明确其转化机制, 为实现水体氮素和 Mn(II) 复合污染物的高效去除提供理论参考。

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